



MPs and NPs intake and heavy metals accumulation in tissues of *Palinurus elephas* (J.C. Fabricius, 1787), from NW Aegean sea, Greece

Thodoros E. Kampouris^{a,*}, Evdokia Syranidou^b, Petroula Seridou^b, Konstantinos Gagoulis^b, Ioannis E. Batjakas^a, Nicolas Kalogerakis^b

^a Department of Marine Sciences, School of the Environment, University of the Aegean, University Hill 81100, Mytilene, Lesvos Island, Greece

^b School of Chemical and Environmental Engineering, Technical University of Crete, Chania, Greece

ARTICLE INFO

Keywords:

Spiny lobster
Marine pollution
Microplastics
Nanoplastics
Threatened species
Bioindicators
Aegean sea

ABSTRACT

European marine waters are infested with plastic, with an average density of 1 plastic item in every four square-meters. Research relevant to MPs-NPs ingestion by wild decapods in European waters is limited, none of which regards the European spiny lobster. Totally 4102 plastic particles were extracted from the spiny lobster stomach tissues of both sexes. Out of the 63 samples analysed only three (4.8%) of them were found with no plastic particles. The range of number of MPs in stomachs was from 20 to 273 MPs individual⁻¹. The 98.3% were fragments. In total 3833 plastic particles were extracted from the gill tissues of both sexes. MPs were found in all samples (n = 50), 99.2% of the detected particles were fragments. The MPs detected in gills ranged from 11 to 339 MPs individual⁻¹. The DLS method was used in order to evaluate the NPs presence. Nanoplastics were detected in 22.6% of stomachs and in the 48.1% of gills. A total of 43 polymer types were identified in both tissues. Also, our study assessed the accumulation of heavy metals at the edible tail muscle. Certain elements were detected above the EU's Maximum Residue Level, including arsenic. The present results are alarming and the potential human health implications could be serious.

1. Introduction

A new threat to marine environment was acknowledged worldwide, meaning the omnipresent plastic pollution. Inappropriate waste management schemes led to an increase of plastics to the marine environment (Alimi et al., 2018). Besides the visible pieces, there are particles of micro-scale, known as microplastics (MPs), that have been detected everywhere: in marine biota (van Sebille et al., 2015), and coastal sediments (e.g., Karkanorachaki et al., 2018; Mistri et al., 2018). Plastic pollution is a rather important marine stressor since it has the potential to harm directly, or by sorption, accumulation, and transfer the marine ecosystem (Nguyen et al., 2019).

European marine waters are infested with plastic, with an average density of 1 plastic item in every four square-meters (Cózar et al., 2015). MPs in Mediterranean waters demonstrate heterogeneity in both distribution and composition, even in sub-basins (Cincinelli et al., 2019). Polyethylene (PE) and polypropylene (PP) particles seem to be the most abundant (Suaria et al., 2016). Plastic particles are ingested by various marine taxa such as the fin whale (*Balaenoptera physalus*) (e.g., Fossi

et al., 2014), the basking shark (*Cetorhinus maximus*) (Fossi et al., 2014), sea turtles (Darmon et al., 2017), large pelagic fish (Romeo et al., 2015), seabirds (Codina-García et al., 2013), and zooplankton (e.g., Panti et al., 2015).

Research relevant to MPs-NPs ingestion by wild decapods in European marine waters is limited. The published data regard the Norway lobster (Murray and Cowie, 2011; Devriese et al., 2017; Cau et al., 2019; Martinelli et al., 2021), the brown shrimp (Devriese et al., 2015), the Symi shrimp (Bordbar et al., 2018) and the blue and red shrimp (Carreras-Colom et al., 2018; Cau et al., 2019), the green crab (Acar and Ateş, 2018; Piarulli et al., 2019) and *Maja squinado* (Welden et al., 2018), none of which concerns the European spiny lobster. Table S1 provides a summary of the current knowledge with respect to the MPs intake in Mediterranean decapod species.

It is documented that nanoplastics (NPs) impact fishes in various ways, since nanoplastics accumulate in tissues, and affect their ethology and growth rates (Barría et al., 2019). Adverse effects of nanoplastics have been also reported on filter feeders such as mussels and oysters. NPs affect the mussels' feeding behaviour (Wegner et al., 2012) and they

* Corresponding author.

E-mail addresses: mard16012@marine.aegean.gr, thodocean@gmail.com (T.E. Kampouris).

could be transferred at the mussels' hemolymph (Sendra et al., 2020). Furthermore, micro- and nano-plastic particles had been observed at oyster larvae (Cole and Galloway, 2015).

Recent studies demonstrated that several crustacean groups, such as larval stages of brine shrimps, barnacles, copepods, and the Antarctic krill could be impacted by various NPs (Ferreira et al., 2019; Gonçalves and Bebianno, 2021). Moreover, NPs diversely affect, at least under laboratory conditions, the condition of the whiteleg shrimp (Chae et al., 2019).

Generally, large sized decapods like crabs, prawns, and crayfishes are considered as bioindicators for the habitats that dwell. For instance, suchlike species are the crab *Callinectes sapidus* Rathbun, 1896 (Salvat-Leal et al., 2020), the prawn *Aristaeomorpha foliacea* (Risso, 1827 in [Risso, 1826–1827]) (Soultani et al., 2021) and the crayfish *Procambarus clarkii* (Girard, 1852) (Goretti et al., 2016). Current research on the interactions between plastics and mercury showed that Hg demonstrated a preference for certain polymer types like PUR and PVC (Santos-Echeandía et al., 2020). Moreover, the exposure route of heavy metals may influence the accumulation rate and the infected tissue (Rivera-Hernández et al., 2019).

A review of recent studies (2015–2021) concerning the accumulation of heavy metals in decapod species from the Mediterranean region was conducted, in order to have some reference points since data regarding the studied species are absent. Recent studies make emphasis on commercially important species due to human consumption and potential health issues. The study Olgunoglu et al. (2015), dealing with *A. foliacea*, reported that the heavy metals concentrations were lower than the limits set by WHO (World Health Organization) and FDA (Food and Drug Administration, USA). Also, Soultani et al. (2021) reported relatively high concentrations of arsenic (As) at *A. foliacea* samples from Ionian Sea. In Mersin Bay, Turkey the concentrations of heavy metals in *Penaeus kerathurus* (Forskål, 1775) were lower than EU's Maximum Residue Level (Korkmaz et al., 2019), see Table S4. A study from Sicily Island, Italy showed high concentrations of mercury (Hg) in the edible tissues of *Parapenaeus longirostris* (H. Lucas, 1846) (Traina et al., 2019). In Israel, the concentration of arsenic in decapods *Marsupenaeus japonicus* and *Portunus pelagicus* were higher, when compared with fish species from the same area and from the same sampling periods (Ramon et al., 2021). The accumulation of heavy metals at the edible muscles of *P. segnis* in the Gulf of Gabes, Tunisia found higher than WHO limits (Annabi et al., 2018). In a study from SE Spain, Segura River the concentrations of lead and cadmium at the edible muscles of the blue crab *C. sapidus* were relatively low (Salvat-Leal et al., 2020). Lastly, in *Squilla mantis* (Linnaeus, 1758), among other studied species, off Tuscany, Italy, the concentrations of cadmium and copper had an evident presence (Bonsignore et al., 2018).

Palinurus elephas (JC Fabricius, 1787), broadly known as European spiny lobster, can be found across the Mediterranean Sea, to the east Atlantic from North Africa to Scotland. It dwells in shallow waters to 200 m (Goñi and Latrouite, 2005; Groeneveld et al., 2013). The European spiny lobster is considered as an omnivorous and scavenging species (Goñi et al., 2001). It is a data-limited species (Marengo et al., 2020), especially for the stocks from east Mediterranean Sea.

The European spiny lobster is classified as "Vulnerable", by the International Union for Conservation of Nature (IUCN), mainly due to its continuous overfishing. It is a highly esteemed and nutritious seafood (Kampouris et al., 2021a), and of high commercial value, reaching a retail price of 90 euros kg⁻¹ (Kampouris et al., 2020). (Kampouris et al., 2020).

The aim of this study was to investigate the MPs ingestion in the European spiny lobster sampled from the coastal waters within the area of the National Marine Park of Alonissos, Northern Sporades. The current study dealt with the occurrence and the abundance of MPs in two tissues of *P. elephas*. This study was the first ever for the species in European-Mediterranean level. Also, the present study assessed the intake of MPs in gill samples of the spiny lobster, being among the very

few -to the best of our knowledge at least, to present suchlike data. Furthermore, the present research dealt with the occurrence and the abundance of NPs, by the use of the DLS (Dynamic Light Scattering) methodology (e.g., Murdock et al., 2008), being the first for Greece and among the few globally. Finally, this study presents data on the accumulation of heavy metals at the edible tissue of the lobsters' tail muscles.

2. Materials & methods

2.1. Sampling methodology

Spiny lobsters were caught, by legally operating coastal fisheries vessels, within the National Marine Park of Alonissos Northern Sporades (NMPANS) and particularly in the coastal waters of Psathoura, Gioura and Kyra Panagia islands. The isles of Psathoura (Zone A2), Gioura (Zone A3), and Kyra Panagia (Zone A4) are within the area of the NMPANS that professional coastal fishing is permitted (NMPANS, 2021) (Fig. S1). Trammel and tangle nets 3–5 km in length, were set over rocky substrate using a 9.5 m-long fishing vessel, at a depth ranging between 40 and 100 m. Measurements of the carapace length, total length and wet weight were obtained, following Kampouris et al. (2020).

2.2. Laboratory procedures

2.2.1. Dissection

All animals were transported live and were euthanised by chilling as described by the protocol of Butler IV (2017). The dissection of the lobster specimens was conducted after Kampouris et al. (2021b), excluding the final stage, meaning the opening of the stomach. The gills of each specimen were isolated right after the removal of the carapace, see Kampouris et al. (2021b) for further details.

2.2.2. Microplastics (MPs) extraction and characterization

The used protocol followed the principles of Avio et al. (2015) and Cau et al. (2019) with specific modifications and adaptations. After dissection, tissue samples (gills and stomachs) were stored at –20 °C, till further analysis. The tissue samples were defrosted at room temperature and were dried at 60 °C for 24 h, then they were potted with distilled water before digestion and extraction. The soft tissues were digested by 20% v/v H₂O₂ solution at 50 °C for 24 h, under continuous shaking. The stomach samples with excess material (food items), were further digested for a couple of hours depending on the load (maximum 24 h). In case the stomachs were full of sand, a density separation step using saturated NaCl solution (density 1.2 g/cm³) was followed (Scott et al., 2019). After digestion the solutions were filtered onto glass filters (Whatman GF/C, pore size: 1.6 µm). The underlying solution of the 50% of the total sample was kept for nanoplastic quantification analysis. The filters were placed in oven for 30 min at 30 °C and were let to cool and were subsequently dyed with the Nile red stain (5 mg/ml) to facilitate MPs detection.

QA/QC protocols were strictly followed to reduce cross-contamination of samples. Briefly, were the following: (i) The surfaces were rigorously washed with distilled water. (ii) The glass equipment was washed with 5% HNO₃ and distilled water, (iii) all treatments were performed in a dedicated laboratory room under a hood, (iv) all main researchers wore cotton made lab robes and clothes, please see Piarulli et al. (2019) for further details. Experimental blanks were tested during extractions and filtration processes.

The retrieved suspected MPs were observed under a dissection stereo-microscope (LEICA MZ7.5), counted, photographed, and categorized according to shape based on the criteria described by Cau et al. (2019). The size of the suspected MPs was also measured at the largest dimension. The analysis of the photos was performed using the software ImageJ version 1.52a. The minimum acceptable size was set to 20 µm in line with the limitation accuracy of the µ-FTIR. A microscopic analysis was also performed under a 10X fluorescence microscope (LEICA) for a

restricted number of filters to identify the suspected MPs from other particles.

The μ -FTIR Nicolet iN10 Thermo Scientific was used for the identification of the polymer type of the MPs. A 10% of the total microplastics counted in stomachs and gills, was analysed (Scott et al., 2019) as follows: 128 scans, a resolution of 4 cm^{-1} , spectral range $4000\text{--}600\text{ cm}^{-1}$. The identification of polymers was performed using the open-source spectral library OpenSpecy (<http://www.openspecy.org>). The matches with Pearson correlation coefficient higher than 0.6 were considered for our analysis (Cowger et al., 2021).

2.2.3. Dynamic Light Scattering (DLS) and nanoplastics (NPs) quantification

The NPs size distribution was measured by the SALD-7500 particle size distribution analyser, (Shimadzu, Japan), which uses DLS method within a range particle diameter between 7 nm and $800\text{ }\mu\text{m}$. The refractive index of material was chosen to be 1.4 which corresponds to polystyrene particles. Three measurements were taken for each tissue, providing information about the (a) number and (b) the volume distribution of the suspected plastic particles.

2.2.4. Heavy metals concentration

The tissues of the tail muscle were placed in an oven at $60\text{ }^{\circ}\text{C}$, till reaching constant weight. When fully dried, the samples (0.3 g) were homogenized using a porcelain mortar by crushing (Simonetti et al., 2013). Tissues were digested, at room temperature with 10 ml of HNO_3 for 2 h. A microwave digestion at $95\text{ }^{\circ}\text{C}$ followed, according to standard procedures (Belabed et al., 2013). After digestion, the mixture was cooled, filtered and ultrapure water was added to reach the volume of 25 ml. Concentrations of heavy metals in lobster tissues were determined by inductively Coupled Plasma Mass Spectrometry (ICP-MS 7500cx coupled with Autosampler Series 3000, both of Agilent Technologies).

The description of Heavy metals QA/QC protocol follows. Heavy metal analysis of the tissues was conducted by an accredited Laboratory at the Technical University of Crete, utilizing an Agilent 7900 Inductively Coupled Plasma Mass Spectrometer (ICPMS). The accredited method is based on US EPA Method 6020 ICPMS including a method blank, control reference, certified reference material (for quality assurance and quality control, blanks and Standard Reference Materials (SRM) were analysed after every 10 sample injections) and a duplicate of each sample matrix in each digestion batch. For the heavy metals profile run, Table S4 includes the LOQ (limit of quantification) and LOD (method detectable limits) of the assay.

2.3. Statistical analyses

Several statistical tests were used to assess the impact of both numerical and factorial variances on the number of MPs in spiny lobsters. The normality of samples was assessed by a Shapiro-Wilk test and homogeneity by a Bartlett test (Kampouris et al., 2021a). Comparison of MPs and NPs presence between sexes and tissues was performed with a Welch's ANOVA test since the samples were not normal and not homogenous. Furthermore, the presence of MPs and NPs in both gill and stomach tissues were analysed pairwise in terms of sex, size (carapace length and total weight) by Student's t-test. In all analyses the significance level of $P=0.05$ was used (Martinelli et al., 2021).

3. Results

3.1. MPs in stomach tissues

We extracted 4102 plastic particles from the tissues of both sexes, 2325 from males and 1705 from females. The number of MPs detected in males is not statistically different when compared to females ($P > 0.05$). Out of the 63 samples analysed only the three (4.8%) were found with no

plastic particles in their stomachs. The range of number of MPs in stomachs was from 20 to 273 MPs individual⁻¹, while the average number of MPs was 76.6 ± 51.5 MPs individual⁻¹. In males the minimum, maximum and average (\pm SD) of MPs was 36, 273 and 79.6 ± 29.2 MPs individual⁻¹ accordingly. In females MPs concentration ranged from 20 to 261 MPs individual⁻¹ with an average concentration of 71.2 ± 59 MPs individual⁻¹. The complete range of MPs occurrence and frequency in the stomach tissues for both male and female individuals is presented at Fig. S2A. The smallest particles have size less than 0.03 mm and the largest ones over 3 mm (Fig. S2B), while the vast majority (88.1%) of particles belong to the 0.02–0.22 mm size class. The 98.3% were fragments and the rest (1.7%) were fibres. No statistical differences were found between the MPs size range in female and male spiny lobsters ($P > 0.05$). All particles smaller than 0.020 mm were excluded from further analysis since they pose some uncertainties on the measurements (i.e., Martinelli et al., 2021).

The relation of individual's weight (W), carapace length (CL), stomach weight (W_s) and stomach fullness with the MPs presence in stomachs was further investigated.

None of the above factors is statistically significant (Welch's ANOVA and Games-Howell post-hoc test) in both female and male individuals ($P < 0.05$). The spiny lobster samples were separated in two categories considering their weight: "small", individuals with weight 85–450 g, and "large" individuals of 451–929 g and in two categories in respect to their Carapace Length (CL): "small", individuals with CL 53–79 mm, and "large" individuals with CL 80–113 mm.

In respect to their weight (W), twenty-four spiny lobsters were characterized as "small" and the minimum, maximum and average \pm SD MPs concentration were 28, 261 and 87.5 ± 58 MPs individual⁻¹ respectively for both male and female individuals. In "small" males the mean MPs concentration was 82.4 ± 16.7 (average \pm SD), while in "small" female lobsters the mean concentration was 105.1 ± 92.7 (average \pm SD). No statistical differences were observed between "small" females and "small" males.

The minimum, maximum, and average values for the "large" individuals ($n = 30$) was 20, 273 and 66.2 ± 44.3 MPs individual⁻¹ respectively. In "large" females the mean MPs concentration was 105.1 ± 92.7 (average \pm SD), while in "large" male lobsters the mean concentration was 80.5 ± 78.7 (average \pm SD). No significant differences were observed between sexes (Fig. S3A).

Eighteen spiny lobsters belonged to the "small" group and thirty-six spiny lobsters were characterized as "large" according to carapace length (CL). The average concentration of MPs were 92.1 ± 65.5 MPs individual⁻¹ and 67.3 ± 23 MPs individual⁻¹ in "small" and "large" group respectively. Specifically, the MPs concentration in "small" individuals ranged from 28 to 261 MPs individual⁻¹ while it ranged from 20 to 273 in "large" individuals.

(Fig. S3B). Significant differences were not observed between the "small" and "large" groups either for male or female spiny lobsters.

Furthermore, all samples were categorized in three groups, regarding their stomach weight. Three groups were formed, samples with stomach weight 1–2.5 g, samples with weight 2.6–5 g and 5.1–13.7 g. Fifteen samples were falling at the first group while the minimum, maximum and average MPs per individual were 39, 261 and 79.6 ± 53.9 MPs individual⁻¹ respectively. Twenty-four samples were falling at the second group. The mean MPs abundance was 72.1 ± 52.7 MPs individual⁻¹ ranging from 20 to 239 MPs individual⁻¹ for both the females and males. The third group included 15 spiny lobsters. The minimum, maximum and average values were 29, 273, 73.6 ± 59.8 MPs individual⁻¹ for the females and the males (Fig. S3C). No significant differences were observed amongst groups within male or female spiny lobsters.

Regarding the stomach fullness, female and male individuals with and without food were observed. The minimum, maximum and average MPs per individual of female lobster with empty stomachs ($n = 14$) was 20, 261 and 83.3 ± 72.6 accordingly. Females with full stomachs ($n = 9$) showed minimum, maximum, and average MPs concentration of 28, 99

and 45.7 ± 22.6 per individual. Twenty-one male lobsters had empty stomachs and the minimum, maximum and average MPs concentration was 40, 155, and 72.7 ± 29.1 MPs individual⁻¹. In males with full stomach the respective values were 43, 273, 94.8 ± 68.6 MPs individual⁻¹ (Fig. S3D). No significant differences were observed amongst groups either for male or female spiny lobsters.

3.2. MPs in gill tissues

In total 3833 plastic particles were extracted from the gill tissues of both sexes. The number of MPs in males does not statistically differs from the number in females (Welch's ANOVA and Games-Howell post-hoc test, $P < 0.05$). MPs were found in all samples ($n = 50$). In detail, the MPs detected in gills ranged from 11 to 339 MPs individual⁻¹, while the average number of MPs was 82.9 ± 58.6 MPs individual⁻¹. In males ($n = 26$) the minimum, maximum and average (\pm SD) of MPs concentration was 11, 339 and 91.6 ± 70.3 MPs individual⁻¹ accordingly. In females the minimum, maximum and average of MPs concentration was 12, 193 and 70.7 ± 41.2 MPs individual⁻¹ respectively. The complete range of MPs occurrence and frequency at the gill tissues for both male and female individuals is presented at Fig. S4A. The smallest particles were 0.02 mm and the largest ones had size over 3 mm, while the size of the majority (67.4%) was 0.02–0.12 mm (Fig. S4B). The 99.2% of the detected particles were fragments and the rest (0.83%) were fibres.

The MPs presence in gill samples, was further analysed, considering the individual's weight (W), carapace length (CL) and gill weight (W_g). The spiny lobster samples were separated into two categories considering their weight and two categories with respect to Carapace Length (CL) as already described above.

Twenty-two "small" lobsters were recorded, considering their total weight. The minimum, maximum, and average \pm SD of MPs values were 11, 339 and 102.1 ± 69.8 MPs per individual. Twenty-eight samples were grouped as "large" (451–929 g), the minimum, maximum, and average \pm SD values for both females and males were 12, 188 and 77.2 ± 47.6 MPs individual⁻¹ (Fig. S5A). Based on the present findings, it seems that the individual's total weight (W) does not significantly affect the MPs presence and abundance in the gills. Thirteen individuals were characterized as "small", based on the CL. The minimum, maximum and average \pm SD of MPs concentration in this group were 41, 339 and 108.9 ± 76.7 MPs individual⁻¹. Thirty-seven spiny lobsters were characterized as "large", and the MPs values (minimum-maximum and average) were 11, 193 and 72.2 ± 47.7 MPs individual⁻¹ (Fig. S5B). Based on the present findings, it seems that the individual's total Carapace Length (CL) does significantly affect the MPs presence and abundance in gills.

The gill samples were further grouped in relation to the gill weight (W_g), and two categories were formed. Samples with gill weight of 2.21–5 g, and samples with gill weight of 5.1–11.7 g. Thirty-three individuals belonged to the first group and seventeen samples were falling at the second group. The average \pm SD number of MPs detected in gills were 83.9 ± 63.1 MPs individual⁻¹ and 77.2 ± 50.3 MPs individual⁻¹ for the first and second group respectively. The MPs abundance ranged from 12 to 339 MPs individual⁻¹ in gills with 2.21–5 g weight and from 11 to 188 in gills with 5.1–11.7 g weight (Fig. S5C). The results of this study suggest that both the gill's weight (W_g) and sex (females-males) are not significant factors. Meaning, that they do not affect the intake rates of MPs in gill tissues.

3.3. NPs in stomach tissues

Nanoplastics were detected in the stomach of 7 out of 31 individuals (22.6%) The minimum, average \pm SD, and maximum size of detected particles in the females were 0.085, 0.331 ± 0.03 and 0.773 μ m, considering measurements per number of particles (Table S2). Accordingly, the values for the males were 0.048, 0.19 ± 0.07 and 0.412 μ m. When considering the volume measurements these values were 0.206, 11.509 ± 4.36 and 33.306 μ m for the females and 0.067, 8.256 ± 3.83

and 47.546 μ m for the males respectively. The distribution of NPs both in the number and volume measurements is illustrated at Figs. S6A and S6B, respectively. The size of NPs in the stomach tissues in female and male spiny lobsters did not statistically differ (Student's t-test and one-way ANOVA test).

3.4. NPs in gill tissues

Nanoplastics were detected in 13 gill samples out of 27 (48.1%). The minimum, average \pm SD, and maximum size of nanoplastics present in females were 0.098, 0.221 ± 0.022 and 0.646 μ m respectively, considering measurements per number of particles. Accordingly, the values for the males were 0.099, 0.277 ± 0.026 and 0.32 μ m. When considering the volume measurements these values for the females were 0.199, 8.997 ± 2.764 and 40.063 μ m. Accordingly, the male spiny lobster samples presented the following values 0.156, 1.374 ± 0.826 and 8.948 μ m. The distribution of NPs both in the number and volume measurements is illustrated at Figs. S6C and S6D, respectively. Table S3 provides the minimum, average \pm SD, and maximum values for all samples and for both measurements (number and volume). The size of NPs in the gill tissues in female and male spiny lobsters did not statistically differ (Student's t-test and one-way ANOVA test).

3.5. Polymer types in MPs

A total of 43 polymer types were identified in the stomachs and gills of the spiny lobster. Representative particles with their reference spectra are presented in Figs. 1 and 2. A more diverse polymer composition (33 different polymer types) was detected in the stomachs compared to the polymer composition in gills (21 different polymer types). The most dominant types were PET, PVC, HDPE and aramid in both tissues (Fig. 3) which accounted for 61% and 63% of the total polymer types identified in stomachs and gills respectively (Fig. 3). The two tissues share 8 polymer types (polycarbonate, PVC, poly (phenylene sulfide), HDPE, PET, PP and polyisoprene chlorinate) while 24 and 12 types were only detected in stomachs and gills respectively.

3.6. Heavy metals accumulation

The complete array of heavy metals was assessed. Several elements were below detection level (Li, V, Cr, Mn, Fe, Co, Ni, Y, Cs, Ba) in both female and male spiny lobsters. On the contrary, arsenic (As) was several times fold of the Maximum Residue Level. The mean \pm SD values of heavy metals concentration at the tail tissues of the spiny lobster samples are presented at Table S4. Also, the table provides a summary, marked in bold, of the MRL as set by E.U. and the World Health Organization (W.H.O) along with the Food and Agriculture Organization of the United Nations (F.A.O).

4. Discussion

Marine plastic pollution is a major issue, and it is directly affecting marine life. Less emphasis is given in decapod crustaceans, at least in European-Mediterranean waters, whereupon studies regard only seven species of which only one is a lobster (*Nephrops norvegicus*), and only one regards the Aegean Sea (*Plesionika narval*), see Table S1. Furthermore, most of the previous studies are dealing with the presence of MPs in the species' stomach, however based on very recent findings, it seems that different tissues may accumulate MPs differently (Cau et al., 2020; McGoran et al., 2020; Martinelli et al., 2021).

The MPs frequency of occurrence was very high in both stomach (95.2%) and gill (100%) tissues. In stomach tissues, including the particles of <0.02 mm–0.20 mm, the MPs number per individual varied a lot (20–277 MPs). The average MPs was 78.1 ± 49.7 . When excluding the MPs <0.20 mm, the average MPs per individual is 4 ± 2.8 particles for the females and 5.2 ± 4.4 for the males, being in accordance with

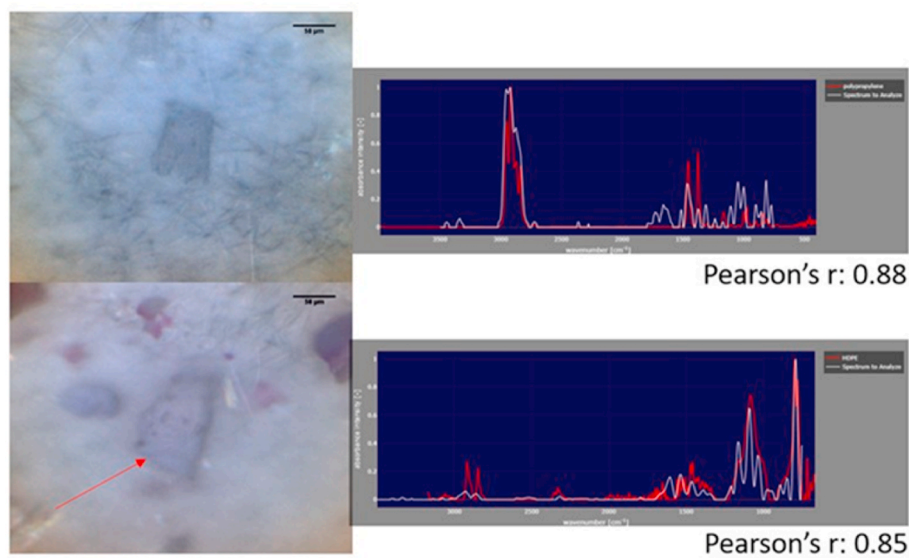


Fig. 1. Representative MPs with their reference spectra, found at the stomach tissues of *P. elephas*.

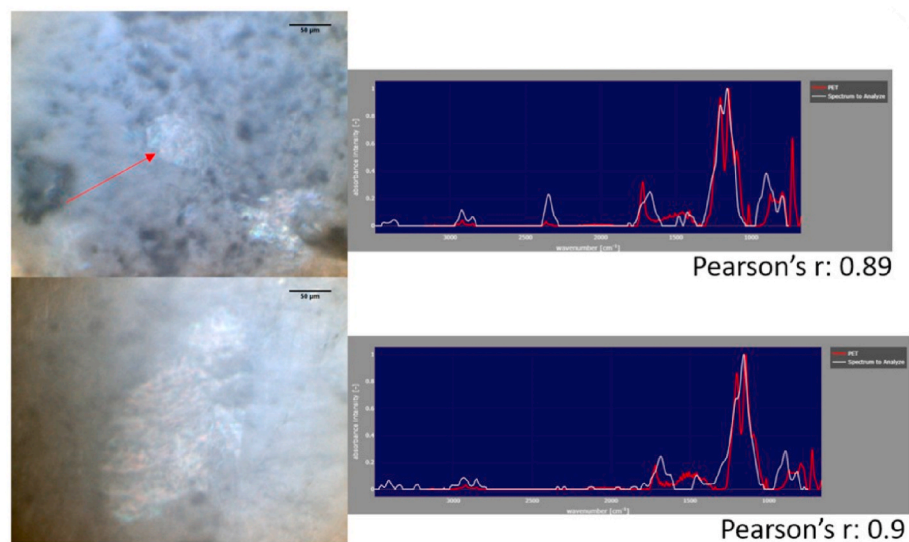


Fig. 2. Representative MPs with their reference spectra, found at the gill tissues of *P. elephas*.

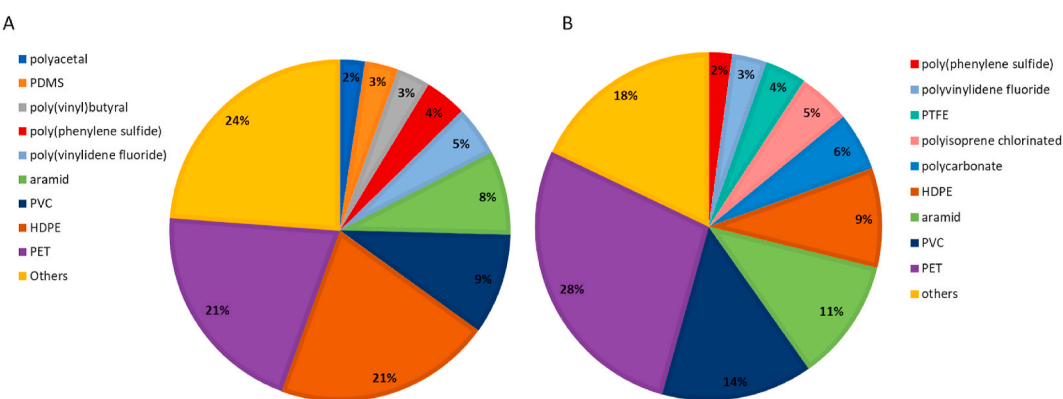


Fig. 3. The polymer types found in stomach (A) and gill (B) samples.

studies regarding the Norway lobster (Cau et al., 2019). The results of this study demonstrated that neither the stomach weight nor the carapace length, nor the total weight of the specimen is a significant factor $P > 0.05$, and do not impact the MPs presence. Meaning, that spiny lobsters -of any size, either with empty stomachs or with full stomachs could ingest similar quantities of MPs. It is worth to mention that in some samples, pieces of fishing nets were found (3–12 cm). Recent studies shown that fishing gear could be a linked source of plastic particles ingestion (Wójcik-Fudalewska et al., 2016; Bordbar et al., 2018).

In gills, including the particles of <0.02 mm– 0.20 mm, the MPs number per individual ranged from 12 to 266 particles, the average was 98.56 ± 59.4 , excluding the <0.20 mm particles the MPs average was 2.9 ± 1.67 per individual, in females and 3 ± 2.52 in males. McGoran et al. (2020), reported very low number (0.2–0.3 in average) of recovered MPs and our results are approximately 10 times fold. The findings of this study showed that neither the total weight nor the carapace length, nor the gills' weight of the specimen is a significant factor $P > 0.05$, and do not impact the MPs presence. However, the MPs presence could be impacted by the weight of the gills and by the active or passive debris removal, a process normally occurring in decapods. Decapods are actively cleaning their gills, in order to remove debris (Baurer, 1999; Batang and Suzuki, 2003). Also, they have another passive way to clean their gills by reversing the respiration mechanism and jetting quantities of water from the gill cavity to the environment (Batang and Suzuki, 1999). Moreover, it is suggested that the feeding strategy could be associated with the MPs ingestion and scavengers such as *P. elephas* (Goñi et al., 2001) may exhibit high rates of MPs, as supported by earlier studies regarding *N. novergicus* (Cau et al., 2019; Martinelli et al., 2021). Based on the above, the scavenging preference of the spiny lobster could perhaps explain the small-sized fibres found at their gills and their origin could be from the suspended sediments, occurring whilst foraging. Watts et al. (2016), demonstrated that, depending on the quantity, plastic microspheres could affect the oxygen consumption when inhaled through the gills, at least for the crab species *Carcinus maenas* (Linnaeus, 1758). The current findings cannot support any similar conclusion for the spiny lobster. The tissues digestion protocol that was used at the current study seems to be accurate and appropriate, since MPs of small size (<0.05 mm) and a few MPs (12 per individual) were found. The protocol used is fulfilling the criteria by Hermesen et al. (2018) and it could be proposed for wider use when processing biological matrices.

Currently, light scattering methodologies, including DLS seem being promising at the detection and quantification of NPs, especially since the absence of standard analytical methods. DLS is proposed to be used in experimental studies that involve organic matrices, like our study (Correia and Loeschner, 2018). Our results suggest that is important to measure the NPs in both per number and per volume since the dominant particles may differ, in a sense that NPs in samples may be smaller and great in number or may be larger in size and less in number.

The results of the current study showed that the main plastic polymer types were PET, HDPE, PVC and Aramid (61% in stomach tissues and 63% in gill tissues). Earlier studies, related with the intake of MPs of *N. novergicus* from Italy demonstrated that the dominant plastic polymers were PES, PA6, PVC and PE (Martinelli et al., 2021). Also, and, Cau et al. (2019) showed that the main MPs types were PE and PP in *N. novergicus* and PE, PP and PES in *A. antenatus* samples. Our results present some similarities with the earlier studies. Contrary, Bodbar et al. (2018), regarding *P. narval*, showed that the dominant MPs were PA types. Studies regarding *C. aestuarii* showed that PES was the dominant polymer type (Piarulli et al., 2019). Furthermore, a recent study from the U.K. demonstrated that PP and PES were dominant in two crab species (McGoran et al., 2020), our study is partially in accordance with the above. The sinking behaviour of MPs in the marine environment is multiparametric and factors like the density of the material, the type (fibre, film, pellet, etc.) along with fouling rate may influence their presence in benthic habitats. Furthermore, materials such as HDPEs are strongly affected by the fouling process (Karkanorachaki et al., 2021),

and that could explain their high abundance at the tissues of the spiny lobster samples.

Our study assessed the accumulation of heavy metals at the edible tail muscle of the spiny lobster. Some elements were below detection limit, see Table S4, however, As was detected above the EU's Maximum Residue Level (MRL) and its concentration was several times fold, suggesting potential human health hazards. As was detected at concentration of 783 times fold of MRL. Earlier studies regarding seawater samples from the region (Ladakis et al., 2003) showed low concentrations of heavy metals. The present results are alarming and the potential human health implications could be serious. Similar studies regarding other spiny lobster species either from the Gulf of California (Morales-Hernández et al., 2004) or from the Persian Gulf (Raissy et al., 2011) showed that only Pb was above specific limits, and their consumption could potentially be a hazard for humans. Also, a study from Bangladeshi waters showed that the concentrations of the edible muscles in spiny lobsters were below the limits set by the Australian authorities (Baki et al., 2018). Squadrone et al. (2021), stated that plastic surfaces can absorb metals, thus acting as vector of metal pollution in marine taxa, increasing the heavy metals concentration in their tissues. Furthermore, the method of exposure could impact both the heavy metals accumulation rate and the impacted tissue. For instance, in mussels the highest Hg accumulation occurred in gills when it was exposed via water. On the contrary, the digestive gland presented higher Hg concentrations via ingestion of plastic particles (Rivera-Hernández et al., 2019). Another potentially important factor, for Hg at least, is the MPs degradation level. It seems that with MPs aging the Hg levels tend to increase (Santos-Echeandía et al., 2020). Noteworthy, is the fact that crustacean species seem to accumulate certain heavy metals like arsenic (As) in higher concentrations when compared to other marine species (Bonsignore et al., 2018; Ramon et al., 2021).

5. Conclusions

Besides the evident ecological and economic value of the spiny lobster, *P. elephas* is an important species for the culture of Mediterranean countries, and it can be said that is amongst the charismatic, flagship species for citizens triggering the environmental awareness and the conservation of threatened species (Kampouris et al., 2022). The results of the present study highlighted that marine pollution is an important stressor that sets additional pressure at the species. Furthermore, considering that the species is consumed by humans, the concentration of certain heavy metals is above the EU's limits and therefore human health hazards arise. Also, the MPs/NPs intake at the spiny lobster tissues stress out, the necessity of continuous and systematic monitoring of the marine pollution effects in a broader geographical scale at the Greek seas.

Considering all the above, meaning that lobsters are highly esteemed, sold in high prices, and that dwell in ecologically important habitats like "maërl" and rocky substrates, see Kampouris et al. (2022) for further details, we propose the adoption of the lobster species as a good bioindicators for MPs and heavy metals pollution and under the EU's Marine Strategy Framework Directive, descriptors 9 and 10.

Author statement

Conceptualization: Thodoros E. Kampouris, Evdokia Syranidou, Nicolas Kalogerakis, Methodology: Thodoros E. Kampouris, Evdokia Syranidou, Petroula Seridou, Validation: Thodoros E. Kampouris, Evdokia Syranidou, Petroula Seridou, Formal analysis: Thodoros E. Kampouris, Investigation: Thodoros E. Kampouris, Evdokia Syranidou, Petroula Seridou, Konstantinos Gagoulis, Resources: Thodoros E. Kampouris, Evdokia Syranidou, Petroula Seridou, Ioannis E. Batjakas, Nicolas Kalogerakis, Data Curation: Thodoros E. Kampouris, Evdokia Syranidou, Petroula Seridou, Konstantinos Gagoulis, Writing - Original Draft: Thodoros E. Kampouris, Evdokia Syranidou, Petroula Seridou,

Ioannis E. Batjakas, Nicolas Kalogerakis, Writing - Review & Editing: Thodoros E. Kampouris, Evdokia Syranidou, Petroula Seridou, Nicolas Kalogerakis, Visualization: Thodoros E. Kampouris, Supervision: Nicolas Kalogerakis.

Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

Data availability

Data will be made available on request.

Appendix A. Supplementary data

Supplementary data to this article can be found online at <https://doi.org/10.1016/j.envpol.2022.120725>.

References

- Acar, S., Ateş, A.S., 2018. Presence of microplastics in the stomachs of *Carcinus aestuarii* nardo, 1857 in çardak lagoon, çanakkale strait, Turkey. *Cah. Biol. Mar.* 59, 493–496. <https://doi.org/10.21411/CBM.A.B1FAB4DA>.
- Alimi, O.S., Farnier Budarz, J., Hernandez, L.M., Tufenkji, N., 2018. Microplastics and nanoplastics in aquatic environments: aggregation, deposition, and enhanced contaminant transport. *Environ. Sci. Technol.* 52 (4), 1704–1724. <https://doi.org/10.1021/acs.est.7b05559>.
- Annabi, A., Bardelli, R., Vizzini, S., Mancinelli, G., 2018. Baseline assessment of heavy metals content and trophic position of the invasive blue swimming crab *Portunus segnis* (Forskål, 1775) in the Gulf of Gabès (Tunisia). *Mar. Pollut. Bull.* 136, 454–463. <https://doi.org/10.1016/j.marpolbul.2018.09.037>.
- Avio, C.G., Gorbi, S., Regoli, F., 2015. Experimental development of a new protocol for extraction and characterization of microplastics in fish tissues: first observations in commercial species from Adriatic Sea. *Mar. Environ. Res.* 111, 18e26. <https://doi.org/10.1016/j.marenvres.2015.06.014>.
- Baki, M.A., Md Hossain, M., Akter, J., Quraishi, S.B., Md Shojib, F.H., Ullah, A.K.M.A., Md Khan, F., 2018. Concentration of heavy metals in seafood (fishes, shrimp, lobster and crabs) and human health assessment in Saint Martin Island, Bangladesh. *Ecotoxicol. Environ. Saf.* 159, 153–163. <https://doi.org/10.1016/j.ecoenv.2018.04.035>.
- Barriá, C., Brands, I., Tort, L., Oliveira, M., Teles, M., 2019. Effect of nanoplastics on fish health and performance: a review. *Mar. Pollut. Bull.* 151, 110791 <https://doi.org/10.1016/j.marpolbul.2019.110791>.
- Batang, Z.B., Suzuki, H., 1999. Gill-cleaning mechanisms of the mud lobster *Thalassinia anomala* (Decapoda: thalassinidea: Thalassinidae). *J. Crustac. Biol.* 19 (4), 671–683. <https://doi.org/10.2307/1549290>.
- Batang, Z.B., Suzuki, H., 2003. Gill-cleaning mechanisms of the burrowing thalassinidean shrimps *Nihonotrypaea japonica* and *Upogebia major* (Crustacea: Decapoda). *J. Zool.* 261 (1), 69–77. <https://doi.org/10.1017/S0952836903003959>.
- Baurer, R.T., 1999. Gill-cleaning mechanisms of a dendrobranchiate shrimp, (Decapoda, penaeidae): description and experimental testing of function. *J. Morphol.* 242, 125–139. [https://doi.org/10.1002/\(SICI\)1097-4687\(199911\)242:2<125::AID-JMOR5>3.0.CO;2-O](https://doi.org/10.1002/(SICI)1097-4687(199911)242:2<125::AID-JMOR5>3.0.CO;2-O).
- Belabed, B.-E., Laffray, X., Dhib, A., Fertouna-Belakhal, M., Turki, S., Aleya, L., 2013. Factors contributing to heavy metal accumulation in sediments and in the intertidal mussel *Perna perna* in the Gulf of Annaba (Algeria). *Mar. Pollut. Bull.* 74, 477–489. <https://doi.org/10.1016/j.marpolbul.2013.06.004>.
- Bonsignore, M., Salvaggio Manta, D., Minto, S., Quinci, E.A., Ape, F., Montalto, V., Gristina, M., Traina, A., Sprovieri, M., 2018. Bioaccumulation of heavy metals in fish, crustaceans, molluscs and echinoderms from the Tuscany coast. *Ecotoxicol. Environ. Saf.* 162, 554–562. <https://doi.org/10.1016/j.ecoenv.2018.07.044>.
- Bordbar, L., Kapisir, K., Kalogirou, S., Anastasopoulou, A., 2018. First evidence of ingested plastics by a high commercial shrimp species (*Plesionika narval*) in the eastern Mediterranean. *Mar. Pollut. Bull.* 136, 472–476. <https://doi.org/10.1016/j.marpolbul.2018.09.030>.
- Butler IV, M.J., 2017. Collecting and processing lobsters. *J. Crustac. Biol.* 37, 340–346. <https://doi.org/10.1093/jcbl/rux021>.
- Carreras-Colom, E., Constenla, M., Soler-Membrives, A., Cartes, J.E., Baeza, M., Padrós, F., Carrasón, M., 2018. Spatial occurrence and effects of microplastic ingestion on the deep-water shrimp *Aristeus antennatus*. *Mar. Pollut. Bull.* 133, 44–52. <https://doi.org/10.1016/j.marpolbul.2018.05.012>.
- Cau, A., Avio, C.G., Dessi, C., Follesa, M.C., Moccia, D., Regoli, F., Pusceddu, A., 2019. Microplastics in the crustaceans *Nephrops norvegicus* and *Aristeus antennatus*: Flagship species for deep-sea environments? *Environ. Pollut.* 255, 113107 <https://doi.org/10.1016/j.envpol.2019.113107>.
- Cau, A., Avio, C.G., Dessi, C., Moccia, D., Pusceddu, A., Regoli, F., Cannas, R., Follesa, M. C., 2020. Benthic Crustacean digestion can modulate the environmental fate of microplastics in the deep sea. *Environ. Sci. Technol.* 54, 4886–4892. <https://doi.org/10.1021/acs.est.9b07705>.
- Chae, Y., Kim, D., Choi, M.-J., Cho, Y., An, Y.-J., 2019. Impact of nano-sized plastic on the nutritional value and gut microbiota of whiteleg shrimp *Litopenaeus vannamei* via dietary exposure. *Environ. Int.* 130, 104848 <https://doi.org/10.1016/j.envint.2019.05.042>.
- Cincinelli, A., Martellini, T., Guerranti, C., Scopetani, C., Chelazzi, D., Giarrizzo, T., 2019. A potpourri of microplastics in the sea surface and water column of the Mediterranean Sea. *Trends Anal. Chem.* 110, 321–326. <https://doi.org/10.1016/j.trac.2018.10.026>.
- Codina-García, M., Militão, T., Moreno, J., González-Solís, J., 2013. Plastic debris in Mediterranean seabirds. *Mar. Pollut. Bull.* 77, 220–226. <https://doi.org/10.1016/j.marpolbul.2013.10.002>.
- Cole, M., Galloway, T.S., 2015. Ingestion of nanoplastics and microplastics by Pacific oyster larvae. *Environ. Sci. Technol.* 49 (24), 14625–14632. <https://doi.org/10.1021/acs.est.5b04099>.
- Correia, M., Loeschner, K., 2018. Detection of nanoplastics in food by asymmetric flow field-flow fractionation coupled to multi-angle light scattering: possibilities, challenges and analytical limitations. *Anal. Bioanal. Chem.* 410, 5603–5615. <https://doi.org/10.1007/s00216-018-0919-8>.
- Cowger, W., Steinmetz, Z., Gray, A., Munno, K., Lynch, J., Hapich, H., Primpke, S., De Frond, H., Rochman, C., Herodotou, O., 2021. Microplastic spectral classification needs an open source community: open specy to the rescue. *Anal. Chem.* 93, 7543–7548. <https://doi.org/10.1021/acs.analchem.1c00123>.
- Cózar, A., Sanz-Martín, M., Martí, E., González-Gordillo, I., Ubeda, B., Gálvez, J.Á., Irigoien, X., Duarte, C.M., 2015. Plastic accumulation in the Mediterranean Sea. *PLoS One* 10 (4), e0121762. <https://doi.org/10.1371/journal.pone.0121762>.
- Darmon, G., Miaud, C., Claro, F., Doremus, G., Galgani, F., 2017. Risk assessment reveals high exposure of sea turtles to marine debris in French Mediterranean and metropolitan Atlantic waters. *Deep Sea Res. Part II* 141, 319–328. <https://doi.org/10.1016/j.dsr2.2016.07.005>.
- Devriese, L.I., De Witte, B., Vethaak, A.D., Hostens, K., Leslie, H.A., 2017. Bioaccumulation of PCBs from microplastics in Norway lobster (*Nephrops norvegicus*): an experimental study. *Chemosphere* 186, 10–16. <https://doi.org/10.1016/j.chemosphere.2017.07.121>.
- Devriese, L.I., van der Meulen, M.D., Maes, T., Bekaert, T., Bekaert, K., Paul-Pont, I., Frère, L., Robbens, J., Vethaak, A.D., 2015. Microplastic contamination in brown shrimp (*Crangon crangon*, Linnaeus 1758) from coastal waters of the southern north sea and channel area. *Mar. Pollut. Bull.* 98, 179–187. <https://doi.org/10.1016/j.marpolbul.2015.06.051>.
- Ferreira, I., Venâncio, C., Lopes, I., Oliveira, M., 2019. Nanoplastics and marine organisms: what has been studied? *Environ. Toxicol. Pharmacol.* 67, 1–7. <https://doi.org/10.1016/j.etap.2019.01.006>.
- Fossi, M.C., Coppola, D., Baini, M., Giannetti, M., Guerranti, C., Marsili, L., Panti, C., de Sabata, E., Clò, S., 2014. Large filter feeding marine organisms as indicators of microplastic in the pelagic environment: the case studies of the Mediterranean basking shark (*Cetorhinus maximus*) and fin whale (*Balaenoptera physalus*). *Mar. Environ. Res.* 100, 17–24. <https://doi.org/10.1016/j.marenvres.2014.02.002>.
- Gonçalves, J.M., Bebianno, M.J., 2021. Nanoplastics impact on marine biota: a review. *Environ. Pollut.* 273, 116426 <https://doi.org/10.1016/j.marenvres.2014.01.001>.
- Goni, R., Latrouite, D., 2005. Review of the biology and fisheries of *Palinurus* spp. species of European waters: *Palinurus elephas* (Fabricius, 1787) and *Palinurus mauritanicus* (Gravel, 1911). *Cah. Biol. Mar.* 46 (2), 127–142. <https://doi.org/10.21411/CBM.A.D8C20491>.
- Goni, R., Quetglas, A., Reñones, O., 2001. Diet of the spiny lobster *Palinurus elephas* (Decapoda: Palinuridae) from the columbretes islands marine reserve (north-western mediterranean). *JMBA (J. Mar. Biol. Assoc.)* 81 (2), 347–348. <https://doi.org/10.1017/S0025315401003861>.
- Goretti, E., Pallottini, M., Ricciarini, M.I., Selvaggi, R., Cappelletti, D., 2016. Heavy metals bioaccumulation in selected tissues of red swamp crayfish: an easy tool for monitoring environmental contamination levels. *Sci. Total Environ.* 559, 339–346. <https://doi.org/10.1016/j.scitotenv.2016.03.169>.
- Groeneveld, J.C., Goni, R., Díaz, D., 2013. *Palinurus* species. In: Phillips, B.F. (Ed.), *Lobsters: Biology, Management, Aquaculture and Fisheries*, second ed. Wiley-Blackwell, Oxford, pp. 326–356.
- Hermesen, E., Mintenig, S.M., Besseling, E., Koelmans, A.A., 2018. Quality criteria for the analysis of microplastic in biota samples: a critical review. *Environ. Sci. Technol.* 52, 10230–10240. <https://doi.org/10.1021/acs.est.8b01611>.
- Kampouris, T.E., Asimaki, A., Klauoudatos, D., Exadactylos, A., Karapanagiotidis, I.T., Batjakas, I.E., 2021a. Nutritional quality of the European spiny lobster *Palinurus elephas* (J.C. Fabricius, 1787) (achelata, palinuridae) and the non-indigenous northern Brown shrimp *Penaeus aztecus* ives, 1891 (dendrobranchiata, penaeidae). *Foods* 10 (10), 2480. <https://doi.org/10.3390/foods10102480>.
- Kampouris, T.E., Gafas, G.A., Sarantopoulou, J., Exadactylos, A., Batjakas, I.E., 2021b. An American in the Aegean: first record of the American lobster *Homarus americanus* H. Milne Edwards, 1837 from the eastern Mediterranean Sea. *Bioinvasions Rec* 10 (1), 170–180. <https://doi.org/10.3391/bir.2021.10.1.18>.
- Kampouris, T.E., Koutsoubas, D., Kanelopoulou, K., Zannaki, K., Batjakas, I.E., 2022. Informing the general public on the threat status of the European spiny lobster, *Palinurus elephas* (Fabricius, 1787) through Citizen-Science and social media platforms. A case study from the Aegean Sea. *Mediterr. Mar. Sci.* 23 (2), 366–373. <https://doi.org/10.12681/mms.26929>.
- Kampouris, T.E., Koutsoubas, D., Milenkova, D., Economidis, G., Tamvakidis, S., Batjakas, I.E., 2020. New data on the biology and fisheries of the threatened *Palinurus elephas* (Fabricius, 1787) (Decapoda, achelata, palinuridae) from the north-west Aegean Sea, Greece. *Water* 12 (9), 2390. <https://doi.org/10.3390/w12092390>.

- Karkanorachaki, K., Kiparissis, S., Kalogerakis, G.C., Yiantzi, E., Psillakis, E., Kalogerakis, N., 2018. Plastic pellets, meso- and microplastics on the coastline of Northern Crete: distribution and organic pollution. *Mar. Pollut. Bull.* 133, 578–589. <https://doi.org/10.1016/j.marpolbul.2018.06.011>.
- Karkanorachaki, K., Syranidou, E., Kalogerakis, N., 2021. Sinking characteristics of microplastics in the marine environment. *Sci. Total Environ.* 793, 148526 <https://doi.org/10.1016/j.scitotenv.2021.148526>.
- Korkmaz, C., Ay, Ö., Çolakfakioğlu, C., Erdem, C., 2019. Heavy metal levels in some edible Crustacean and mollusk species marketed in mersin. *Thalassas* 35, 65–71. <https://doi.org/10.1007/s41208-018-0086-x>.
- Ladakis, M., Skoullou, M., Dassenakis, M., 2003. Water quality in a mediterranean marine protected area (north Sporades islands, Greece). *Chem. Ecol.* 19 (1), 47–57. <https://doi.org/10.1080/0275754031000084400>.
- Marengo, M., Theuerkauff, D., Patrissi, M., Doutreloux, N., Leduc, M., et al., 2020. A typical mediterranean fishery and an iconic species: focus on the common spiny lobster (*Palinurus elephas*, Fabricius, 1787) in corsica. *Oceanogr Fish* 12 (1). <https://doi.org/10.19080/OFOAJ.2020.12.555827>. OFOAJ.MS.ID.5557827.
- Martinelli, M., Gomiero, A., Guicciardi, S., Frapiccini, E., Straffella, P., Angelini, S., Domenichetti, F., Belardinelli, A., Colella, S., 2021. Preliminary results on the occurrence and anatomical distribution of microplastics in wild populations of *Nephrops norvegicus* from the Adriatic Sea. *Environ. Pollut.* 278, 116872 <https://doi.org/10.1016/j.envpol.2021.116872>.
- McGoran, A.R., Clark, P.F., Smith, B.D., Morritt, D., 2020. High prevalence of plastic ingestion by *Eriocheir sinensis* and *Carcinus maenas* (Crustacea: Decapoda: Brachyura) in the Thames Estuary. *Environ. Pollut.* 265, 114972 <https://doi.org/10.1016/j.envpol.2020.114972>.
- Mistri, M., Infantini, V., Scoptoni, M., Granata, T., Moruzzi, L., Massara, F., De Donati, M., Munari, C., 2018. Microplastics in marine sediments in the area of pianosa Island (central adriatic sea). *Rendiconti Lincei. Sci. Fis. Nat.* 29, 805–809. <https://doi.org/10.1007/s12210-018-0736-1>.
- Morales-Hernández, F., Soto-Jiménez, M.F., Páez-Osuna, F., 2004. Heavy metals in sediments and lobster (*Panulirus gracilis*) from the discharge area of the submarine sewage outfall in Mazatlán Bay (SE Gulf of California). *AECT* 46, 485–491. <https://doi.org/10.1007/s00244-003-3064-z>.
- Murdock, R.C., Braydich-Stolle, L., Schrand, A.M., Schlager, J.J., Hussain, S.M., 2008. Characterization of nanomaterial dispersion in solution prior to *in vitro* exposure using dynamic light scattering technique. *Toxicol. Sci.* 101 (2), 239–253. <https://doi.org/10.1093/toxsci/kfm240>.
- Murray, F., Cowie, P.R., 2011. Plastic contamination in the decapod crustacean *Nephrops norvegicus* (Linnaeus, 1758). *Mar. Pollut. Bull.* 62, 1207–1217. <https://doi.org/10.1016/j.marpolbul.2011.03.032>.
- National Marine Park of Alonissos Northern Sporades, 2021. NMPANS). https://alonissos-park.gr/?page_id=2&lang=en_US. (Accessed 25 May 2021). accessed on.
- Nguyen, B., Claveau-Mallet, D., Hernandez, L.M., Genbo Xu, E., Farner, J.M., Tufenkji, N., 2019. Separation and analysis of microplastics and nanoplastics in complex environmental samples. *Acc. Chem. Res.* 52 (4), 858–866. <https://doi.org/10.1021/acs.accounts.8b00602>.
- Olgunoğlu, M.P., Olgunoğlu, İ.A., Bayhan, Y.K., 2015. Heavy metal concentrations (Cd, Pb, Cu, Zn, Fe) in giant red shrimp (*Aristaeomorpha foliacea* Risso 1827) from the Mediterranean Sea. *Pol. J. Environ. Stud.* 24 (2), 631–635. <https://doi.org/10.15244/pjoes/33201>.
- Panti, C., Giannetti, M., Baini, M., Rubegni, F., Minutoli, R., Fossi, M.C., 2015. Occurrence, relative abundance and spatial distribution of microplastics and zooplankton NW of sardinia in the pelagos sanctuary protected area, Mediterranean Sea. *Environ. Chem.* 12 (5), 618–626. <https://doi.org/10.1071/EN14234>.
- Piarulli, S., Scapinello, S., Comandini, P., Magnusson, K., Granberg, M., Wong, J.X.W., Sciutto, G., Prati, S., Mazzeo, R., Booth, A.M., Airolidi, L., 2019. Microplastic in wild populations of the omnivorous crab *Carcinus aestuarii*: a review and a regional-scale test of extraction methods, including microfibrils. *Environ. Pollut.* 251, 117–127. <https://doi.org/10.1016/j.envpol.2019.04.092>.
- Raissy, M., Ansari, M., Rahimi, E., 2011. Mercury, arsenic, cadmium and lead in lobster (*Panulirus homarus*) from the Persian Gulf. *Toxicol. Ind. Health* 27 (7), 655–659. <https://doi.org/10.1177/0748233710395346>.
- Ramon, D., Morick, D., Croot, P., Berzak, R., Scheinin, A., Tchernov, D., Davidovich, N., Britzi, M., 2021. A survey of arsenic, mercury, cadmium, and lead residues in seafood (fish, crustaceans, and cephalopods) from the south-eastern Mediterranean Sea. *J. Food Sci.* 86 (3), 1153–1161. <https://doi.org/10.1111/1750-3841.15627>.
- Rivera-Hernández, J.R., Fernández, B., Santos-Echeandía, J., Garrido, S., Morante, M., Santos, P., Albentosa, M., 2019. Biodynamics of mercury in mussel tissues as a function of exposure pathway: natural vs microplastic routes. *Sci. Total Environ.* 674, 412–423. <https://doi.org/10.1016/j.scitotenv.2019.04.175>.
- Romeo, T., Pietro, B., Pedà, C., Consoli, P., Andaloro, F., Fossi, M.C., 2015. First evidence of presence of plastic debris in stomach of large pelagic fish in the Mediterranean Sea. *Mar. Pollut. Bull.* 95, 358–361. <https://doi.org/10.1016/j.marpolbul.2015.04.048>.
- Salvat-Leal, I., Verdiell, D., Parrondo, P., Barcala, E., 2020. Assessing lead and cadmium pollution at the mouth of the river Segura (SE Spain) using the invasive blue crab (*Callinectes sapidus* Rathbun, 1896, Crustacea, Decapoda, Portunidae) as a bioindicator organism. *Reg. Stud. Mar. Sci.* 40, 101521 <https://doi.org/10.1016/j.rsma.2020.101521>.
- Santos-Echeandía, J., Rivera-Hernández, J.R., Rodrigues, J.P., Moltó, V., 2020. Interaction of mercury with beached plastics with special attention to zonation, degradation status and polymer type. *Mar. Chem.* 222, 103788 <https://doi.org/10.1016/j.marchem.2020.103788>.
- Scott, N., Porter, A., Santillo, D., Simpson, H., Lloyd-Williams, S., Lewis, C., 2019. Particle characteristics of microplastics contaminating the mussel *Mytilus edulis* and their surrounding environments. *Mar. Pollut. Bull.* 146, 125–133. <https://doi.org/10.1016/j.marpolbul.2019.05.041>.
- Sendra, M., Saco, A., Yeste, M.P., Romero, A., Novoa, B., Figueras, A., 2020. Nanoplastics: from tissue accumulation to cell translocation into *Mytilus galloprovincialis* hemocytes. resilience of immune cells exposed to nanoplastics and nanoplastics plus *Vibrio splendidus* combination. *J. Hazard Mater.* 388, 121788 <https://doi.org/10.1016/j.jhazmat.2019.121788>.
- Simonetti, P., Botté, S.E., Fiori, S.M., Marcovecchio, J.E., 2013. Burrowing crab (*Neohelice granulata*) as a potential bioindicator of heavy metals in the bahía blanca estuary, Argentina. *Arch. Environ. Contam. Toxicol.* 64, 110–118. <https://doi.org/10.1007/s00244-012-9804-1>.
- Soultani, G., Sele, V., Rasmussen, R.R., Pasiás, I., Stathopoulou, E., Thomaidis, N.S., Sinanoglu, V.J., Sloth, J.J., 2021. Elements of toxicological concern and the arsenolipids' profile in the giant-red Mediterranean shrimp, *Aristaeomorpha foliacea*. *J. Food Compos. Anal.* 97, 103786 <https://doi.org/10.1016/j.jfca.2020.103786>.
- Squadron, S., Pederiva, S., Bezzo, T., Sartor, R.M., Battuello, M., Nurra, N., Griglione, A., Brizio, P., Abete, M.C., 2021. Microplastics as vectors of metals contamination in Mediterranean Sea. *Environ. Sci. Pollut. Res.* <https://doi.org/10.1007/s11356-021-13662-7>.
- Suaría, G., Avio, C.G., Mineo, A., Lattin, G.L., Magaldi, M.G., Belmonte, G., Moore, C.J., Regoli, F., Aliani, S., 2016. The Mediterranean Plastic Soup: synthetic polymers in Mediterranean surface waters. *Sci. Rep.* 6, 37551 <https://doi.org/10.1038/srep37551>.
- Traina, A., Bono, G., Bonsignore, M., Falco, F., Giuga, M., Quinci, E.A., Vitale, S., Sprovieri, M., 2019. Heavy metals concentrations in some commercially key species from Sicilian coasts (Mediterranean Sea): potential human health risk estimation. *Ecotoxicol. Environ. Saf.* 168, 466–478. <https://doi.org/10.1016/j.ecoenv.2018.10.056>.
- van Sebille, E., Wilcox, C., Lebreton, L., Maximenko, N., Hardesty, B.D., van Franeker, J. A., Eriksen, M., Siegel, D., Galgani, F., Lavender Law, K., 2015. A global inventory of small floating plastic debris. *Environ. Res. Lett.* 10, 124006 <https://doi.org/10.1088/1748-9326/10/12/124006>.
- Watts, A.J.R., Urbina, M.A., Goodhead, R., Moger, J., Lewis, C., Galloway, T.S., 2016. Effect of microplastic on the gills of the shore crab *Carcinus maenas*. *Environ. Sci. Technol.* 50, 5364–5369. <https://doi.org/10.1021/acs.est.6b01187>.
- Wegner, A., Besseling, E., Foekema, E.M., Kamermans, P., Koelmans, A.A., 2012. Effects of nanoplastystyrene on the feeding behavior of the blue mussel (*Mytilus edulis*). *Environ. Toxicol. Chem.* 31 (11), 2490–2497. <https://doi.org/10.1002/etc.1984>.
- Welden, N.A., Abylkhani, B., Howarth, L.M., 2018. The effects of trophic transfer and environmental factors on microplastic uptake by plaice, *Pleuronectes platessa*, and spider crab, *Maja squinado*. *Environ. Pollut.* 239, 351–358. <https://doi.org/10.1016/j.envpol.2018.03.110>.
- Wójcik-Fudalewska, D., Normant-Saremba, M., Anastácio, P., 2016. Occurrence of plastic debris in the stomach of the invasive crab *Eriocheir sinensis*. *Mar. Pollut. Bull.* 113, 306–311. <https://doi.org/10.1016/j.marpolbul.2016.09.059>.